

Dye wastewater treatment by using ozonation/ photocatalytic processes

Farhad Qaderi^{*,1}, Amir Hossein Norouzi²

¹Faculty of Civil Engineering, Babol Noshirvani University of Technology, Babol, Iran.

²Faculty of Technical and Engineering, Payame Noor University (PNU), Tehran, Iran.

GRAPHICAL ABSTRACT



ARTICLE INFO

Article type:

Research Article

Article history:

Received x x x

Received in revised form x x x

Accepted x x x

Available online

Keywords:

Dye wastewater

Ozonation process

Immobilized photocatalyst process

Kinetic

Intermediate products



© The Author(s)

Publisher: Razi University

ABSTRACT

In this research, the capability of two improved photocatalytic (immobilized-suspended) and hybrid (ozonation/Improved photocatalytic) methods were studied to treat wastewater containing azo dye of Red 40. Hybrid process had the best efficiency for dye concentration of 50 mg/L with pH of 9 and TiO₂ concentration of 40 g/m³ with light source power of 120 W and ozone injection rate of 20 mg/min. In these conditions, dye and COD were removed completely after 65 min and 7 h. Based on the kinetic studies, pseudo first order model was applied for dye degradation. According to the results, immobilization of nano materials not only solved their recovery problem, but also improved the efficiency of conventional photocatalytic process. Addition of ozonation to the photocatalytic system for the purpose of having hybrid process was also resulted in an increase of the efficiency to over 30 times and production of safe products. The results have been showed the advantages of hybrid process.

1. Introduction

Papers Treatment of dye wastewater is an important issue in environmental researches, specially using advanced oxidation processes are used in this field recently. Annually, over 7×10⁵ tons of dyes is generated in the world (Gomes, Nunes, and Simões, 2010). Dye wastewaters have usually color and high temperature (Hu, Lin and Chiang, 2004). Azo dyes have been used in the color paper printing, textile, paper making, food, leather and cosmetic industries (Prahba and Lathasree, 2013). Moreover, some azo dyes have been found to

be human carcinogens (Behnajady *et al.*, 2009). These kinds of compounds are classified as environmental hazardous materials, due to their toxicity and hard biodegradation (Lu *et al.*, 2009) which have to be treated before discharging to the environment. Different physical methods such as adsorption (Khani *et al.*, 2013), ultra-filtration (Kasperchik, Yaskevich and Bil'dyukevich, 2010), reverse osmosis (Behnajady *et al.*, 2009) and electro coagulation are widely used for dyes removal but they transport contaminants from water to sludge (Behnajady *et al.*, 2009). Conventional biological treatment methods are ineffective for decolorization and degradation of dyes, because of

*Corresponding author Email: f.qaderi@nit.ac.ir

presence of hard biodegradable aromatics in wastewaters (El-Ashtouky, 2013). Advanced Oxidation Processes (AOPs) such as ozonation and photocatalytic methods are proper to mineralize dye wastewater (Zuorroa and Lavecchia, 2013; Peng *et al.*, 2008; Qaderi and Tamadoni, 2024; Ghasemi, Zoqi and Zanganeh Ranjbar, 2024; Ayati, 2018; Qaderi and Abdolalian, 2022; Farzinfar and Qaderi, 2022). They are characterized by production of hydroxyl radicals (Weng, Zhou and Zhang, 2013). Radicals are extremely reactive oxidants (Moradianfard, Ghorbani and Bakhshy, 2013). Ozone has been seen as an effective agent to remove the color of dye wastewater especially azo dyes (Turhan and Turgut, 2009; Rabieian and Qaderi, 2024; Tamadoni and Qaderi, 2019; Khoureshidi and Qaderi, 2023; Qaderi and Banisheykholeslami, 2025). Ozonation of textile wastewater with 600 mg/L of initial COD resulted in 80 percent efficiency in 140 minutes (Chu *et al.*, 2008). 120 L/hr ozone injection flow with 24 gr/m³ ozone concentration decolorized 2000 mL of 800 mg/L Sirius Blue SBRR during 26 minutes (Turhan and Turgut, 2009), but ozonation is an expensive and unsafe process (Peng *et al.*, 2008).

TiO₂ photocatalytic process is safe and does not require expensive oxidants (Ranjbar, Ayati and Ganjidoust, 2022; Azizi, Ebadi and Qaderi, 2022; Gholamian *et al.*, 2025). It is based on exposure to a photocatalyst (Pratap Reddy *et al.*, 2004; Augugliaro *et al.*, 2004) which can be carried out at moderate temperature and pressure (Barka *et al.*, 2008; Ranjbar, Ayati and Ganjidoust, 2019). TiO₂ nano particles were used in slurry and immobilized manner for wastewater treatment (Zhu, Zhang and Chen, 2010; Delnavaz *et al.*, 2011). It has been reported that 90 percent of 50 mg/L Red 40 (R40) has been removed by modified TiO₂ with Fe and carbon atoms (Wu *et al.*, 2010) and 20 mg/L R40 wastewater decolorized 80% by Bi₁₂TiO₂₀ (Zhu, Zhang and Chen, 2010).

According to the successful application of media in the attached biological systems (Qaderi, Ayati and Ganjidoust, 2011; Delnavaz, Ayati and Ganjidoust, 2010) and necessity of nano-materials separation in slurry systems, before discharging to the environment and lower efficiency of immobilized manner than systems (Delnavaz *et al.*, 2012), a new method must be used.

Two main aims of this study were to apply a new improved photocatalytic system and a combination of ozonation and improved photocatalytic systems to troubleshoot the drawbacks and use the capabilities of both methods.

2. Materials and methods

2.1. Reactor set-up

A schematic diagram of the reactor is presented in Fig. 1. The reactor was a plexiglas cube with internal size of (W, L and H) 10, 15 and 27 cm and effective volume of 2.6 L. TiO₂ nanoparticles with a specific surface area of 20 g/m² were coated on circular plexiglas media (20 mm diameter and 2 mm thickness) by Hec-8 epoxy binder (Azar Resin Chemical Industrial Co.). The coated media were suspended in the reactor by stirrer (IKA, RH basic 2 models). UV-A lamps were used as the irradiation sources.

Ozone was produced from oxygen (99.9% purity) by ozone generator; Arda COG-5S model. The outlet stream from the ozonizer with 1 L/min flow rate, 1.2 g O₃/h and the pressure of 1 bar was injected to the solution from the diffusers. The gas flow rate was measured by air flow meters (Fischer Porter Co.). The temperature was fixed at 22±1°C during all experiments. Unused ozone was injected into 2% KI (Merck) solution, to provide safety conditions.

2.2. Reagents and catalysts

P25 TiO₂ with specific surface area of 55 m²/g and 70 % anatine crystalline form was procured from Degussa Corporation. R40, H₂SO₄, K₂Cr₂O₇, Ag₂SO₄ (Merck), HgSO₄ (DBH Company), Potassium hydrogen phthalate (Hach Company) was used for chemical oxygen demand (COD) measurement. The pH was adjusted by aqueous solutions of HCl (Merck) and NaOH (Merck). The two distilled water was used in all steps of research.

2.3. Analytical methods

The dye concentration was measured using a spectrophotometer at the maximum absorption wavelength of 485 nm (Agilent/Varian Com., USA). COD and pH measurement were accomplished in accordance with 5220-B and 4500-H*B instructions, respectively, presented in the "Standard Methods for the Examination of Water and Wastewater" (APHA, 2005). All laboratory tests were repeated at least three times and relative standard deviation of the measured data was within 2%. Gas chromatography– mass spectrometry (GC-MS) system was used

for identifying intermediate products and kinetic study. GC (Agilent Co., 7890a model) had a Chrompak CP-Sil 8 CB capillary column (50 m× 250 µm× 0.12 µm film thickness) and interfaced to the MS (Agilent Co, 5475c model). The GC column was operated in a temperature programmed mode. It had an initial temperature of 50 °C held for 2 minutes, rise at 290 °C with a 6 °C/min rate, and maintained at that temperature for 10 min.

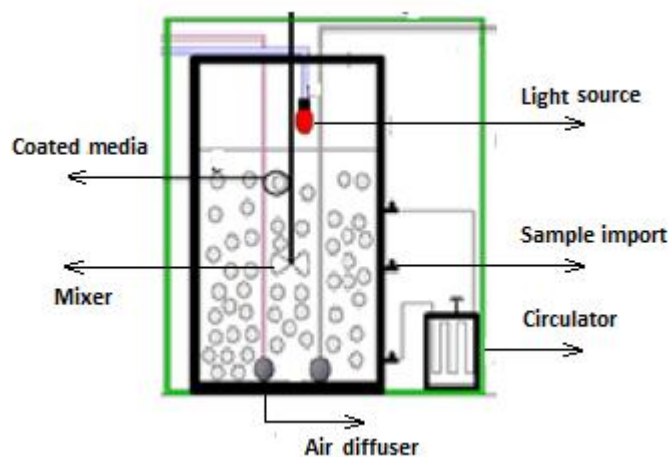


Fig. 1. A schematic diagram of the reactor

2.4. Experimental procedures

In the present research, the photocatalytic and hybrid processes were compared from the stand point of Red 40 removals. The independent variables of dye concentration, pH, and nano-materials concentration and radiation source were investigated. The optimum conditions of both processes were determined (see Table 1). The dye removal intermediates and kinetic considerations were also studied in optimum conditions of hybrid process.

Table 1. Conditions of decolorization experiments.

Variable	Levels
Dye concentration (mg/L)	5, 35, 50
pH	3, 5, 7, 9
Nano TiO ₂ concentration (g/m ²)	20, 40, 60, 80
Irradiation source power (W)	40, 80, 120, 160
Ozone injection rate (mg/min)	12, 20

3. Results and discussion

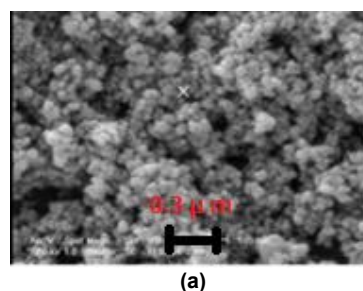
3.1. Immobilized media surface characteristics

The structure and shape of coated TiO₂ is shown in SEM micrographs with magnification of 30000 times (see Fig. 2a). According to the Fig. 2, the coated particles were in nano scale. Energy-dispersive X-ray (EDX) results were also confirmed the presence of TiO₂ on media surface (see Fig. 2b).

This result confirmed that this method of using nano particles in dye wastewater treatment can be used widely and this method is useful for coating of nanoparticles on the surface.

3.2. Effect of solution pH

The dye removal efficiency in different pH conditions are shown in Fig. 3. As shown, when pH increased, the dye removal was increased in photocatalytic and decreased in hybrid system, respectively. Time required to complete the elimination of dye is decreased from 120 minutes in pH=3 to 65 minutes in pH=9. Regarding the higher efficiency of dye removal in acidic pH in photocatalytic process, the hybrid process has a higher efficiency in alkaline pH.



(a)

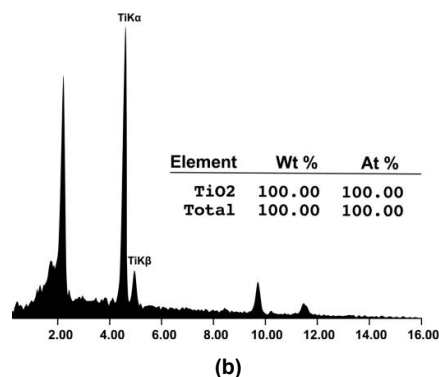


Fig. 2. (a) SEM micrographs (30 kX) and (b) EDX of coated TiO₂ on media.

The pH effect on photocatalytic process efficiency is based on the oxidation capability of produced holes with pH increase and TiO₂ amphoteric behavior (Papadam *et al.*, 2007). According to the literatures, the ozonation and photocatalytic processes have the higher efficiencies in alkaline (Turhan and Turgut, 2009; Somensi *et al.*, 2010) and acidic conditions (Papadam *et al.*, 2007), respectively. Thus, the ozonation is recognized as the dominant process in the hybrid method.

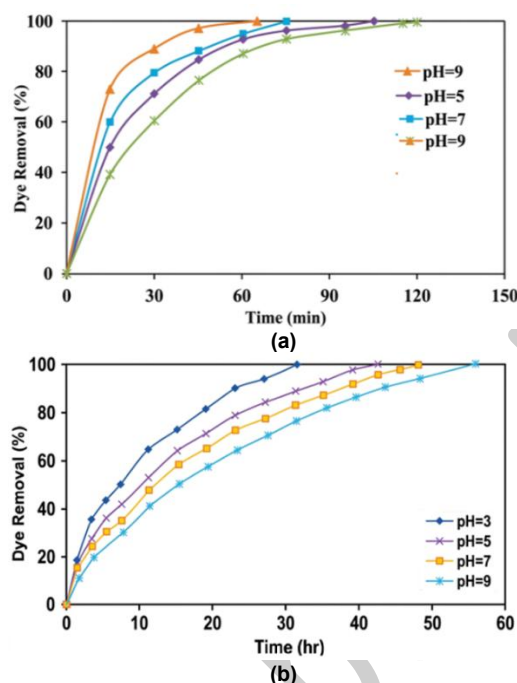


Fig. 3. Dye removal efficiency of (a) hybrid and (b) photocatalytic processes in different pH conditions. ([O₃]=20 mg/L, [R40]=50 mg/L, [TiO₂]=40g/m², P_{UV-A}=120 W).

3.3. Effect of dye concentration

Dye removal for both photocatalytic and hybrid processes in different dye concentrations and optimum pH conditions of 9 and 5 are shown in Fig. 4. Based on the results, this quantity was increased from 3.12 to 3.93 mg/hr when the dye concentration was increased from 5 to 50 mg/L in the photocatalytic process, but it was reached to 3.6 mg/hr as the dye concentration was increased to 60 mg/L. In hybrid process, it was increased from 30 to 115 mg/hr as the dye concentration was increased from 5 to 50 mg/L and reached 102 mg/hr as the initial concentration was increased to 60 mg/L. Because of the hydroxyl radicals' short lifetime, they have a very short time to oxidize the contaminants and the probability of collision with dye is decreased when the dye concentration is reduced.

As it was expected, the dye removal efficiency was decreased when the concentration was increased. For complete removal of 5, 35 and 50 mg/L of dye, the needed time were about 25, 55 and 60 min for hybrid process and 4.17, 22 and 33 hours for photocatalytic process, respectively. As the dye concentration to oxidant ratio is increased, a higher number of bonds should be broken which takes longer time regarding that the oxidant amount is constant. Similar results have been reported in ozonation and photocatalytic processes (Turhan and Turgut, 2009; Barka *et al.*, 2008; Papadam *et al.*, 2007). The removed

dye amounts per unit of time for different concentrations in both systems are compared in Fig. 5.

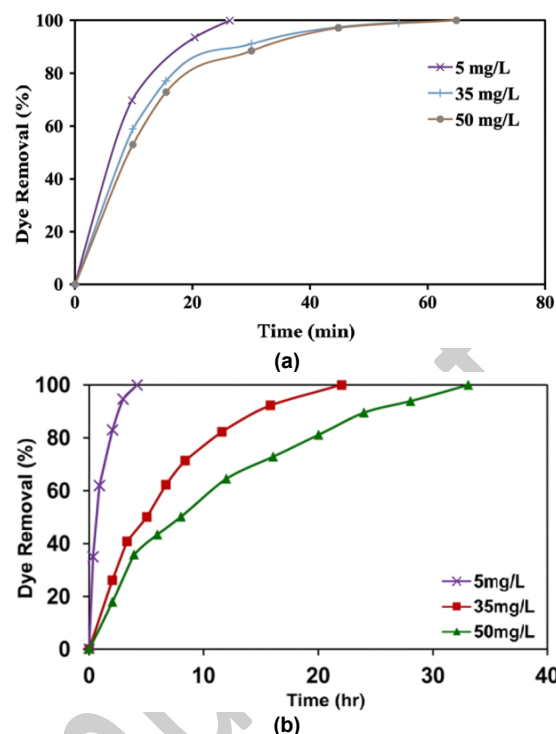


Fig. 4. Dye removal efficiency of a) hybrid and b) photocatalytic processes in different pollutant concentrations ([O₃]=20mg/L, pH_{Hybrid} process=9, pH_{Photocatalytic} process=3, [TiO₂]=40gr/m², P_{UV-A}=120 W).

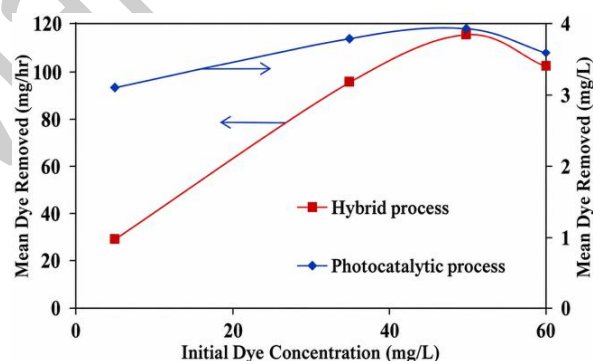


Fig. 5. Initial dye concentration effect on average removed in time unit ([O₃]=20 mg/L, pH_{Hybrid} process=9, pH_{Photocatalytic} process=3, [TiO₂]=40 gr/m², P_{UV-A}=120 W).

This resulted in a decrease of dye removal intensity (Turhan and Turgut, 2009; Barka *et al.*, 2008). Moreover, the intermediate formed concentration was increased as the R40 concentration increased in which these compounds were also consumed the oxidant (Lackey, Mines Jr. and McCreanor, 2006). The interplay between two discussed phenomena specifies the optimum concentration of R40 in dye removal amount per unit of time. According to the results, 50 mg/L was chosen as the optimum concentration in both processes.

3.4. Effect of catalyst concentration

Nano TiO₂ concentration effect on the photocatalytic and hybrid processes efficiency is shown in Fig. 6. Based on the results, the hybrid process had higher efficiency (100 %) in 65 minutes for 40 gr/m² of TiO₂. This efficiency was attributed to an increase in active sites of TiO₂ photocatalyst as the concentration was increased up to 40 gr/m². The similar results have been reported in slurry photocatalytic studies (Kartal, Erol and Oguz, 2001). The enhancement of ozone oxidation in light presence has been reported which justified the higher efficiency of hybrid process as compared to the photocatalytic method (Amat *et al.*, 2005). Based on the obtained results, the amount of 40 gr/m² was chosen as the optimum concentration.

As the nano materials were coated on the media, an increase in the photocatalyst concentration (media's number) was caused UV light penetration into the solution (Kartal, Erol and Oguz, 2001). This matter affects the TiO₂ excitation and synergetic effect of O₃/UV combination

(Amat *et al.*, 2005; Chou and Chang, 2007) that is why the amount of removal efficiency wasn't changed considerably.

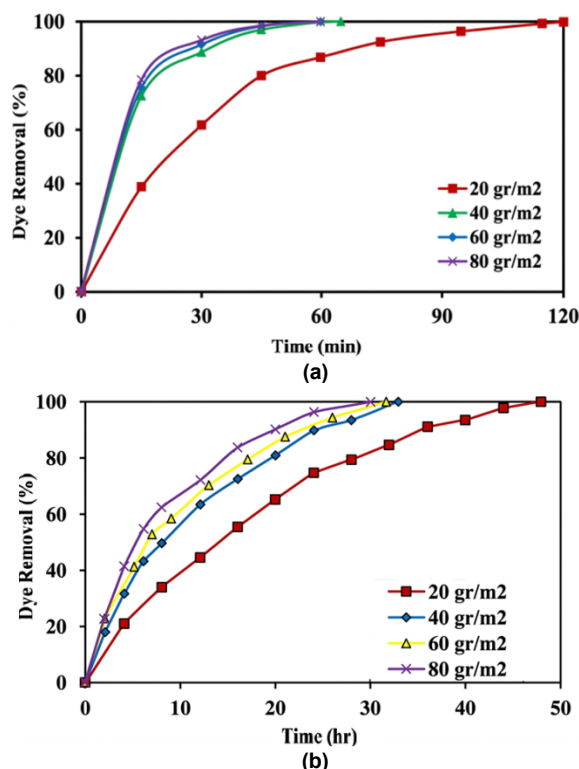
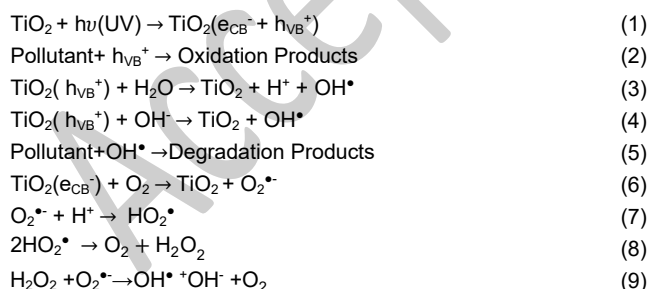


Fig. 6. a) Hybrid and b) photocatalytic dye removal processes efficiency in different nanomaterial concentration ($[O_3]=20\text{mg/L}$, $\text{pH}_{\text{Hybrid process}}=9$, $\text{pH}_{\text{Photocatalytic process}}=3$, $[R40]=50\text{ mg/L}$, $P_{\text{UV-A}}=120\text{ W}$).

3.5. Effect of radiation source power

The effect of different light sources power on the efficiency of hybrid and photocatalytic processes was shown in Fig. 7. It can be seen that the dye removal was increased as the light source power increased in both systems. As the light source power was increased from 40 to 160 Watt, the needed time to complete the dye removal was decreased from 75 to 52 hours and from 100 to 60 minutes in hybrid and photocatalytic processes, respectively. Because in the photocatalytic process, by rising the light source power, rate of the reaction (1) and the electron (e_{CB^-}) and hole (h_{VB^+}) will be increased and the reaction of pollutant oxidation is improved through reactions of (2) to (9) (Fujishima, Rao, and Tryk, 2000; Papadam *et al.*, 2007; Mirkhani *et al.*, 2009).

Also, in the hybrid system, O_3/UV combination had significant synergetic effect on the removal efficiency (Amat *et al.*, 2005; Chou and Chang, 2007).



Based on the results, increase in the efficiency was negligible as the power increased from 120 to 160 Watt due to high dye decomposition rate in high radiation intensities. Because, the light intensity effect of efficiency has a saturation level and no increase in efficiency is obtained in higher intensities (Chiou, Wu, and Juang, 2008). Thus, 120 Watt was applied as the optimum power.

3.6. Comparison of the processes

The comparison of two studied systems is summarized in Table 2. Based on the results, retention time was decreased 30 times as the slurry photocatalytic process was promoted to hybrid and the average

removed dye per unit of time in optimum conditions was reached from 3.8 mg/hr in photocatalytic to 115.7 mg/hr in hybrid method.

Table 2. Comparison optimum condition of improved photocatalytic and hybrid processes.

Parameter	Improved photocatalytic	Hybrid system
pH	3	9
Dye concentration, mg/L	50	50
TiO ₂ concentration, g/m ²	40	40
Light source power, W	120	120
Time of complete dye removal, h	33	1.08
Removed dye in a time unit, mg/h	3.8	115.7

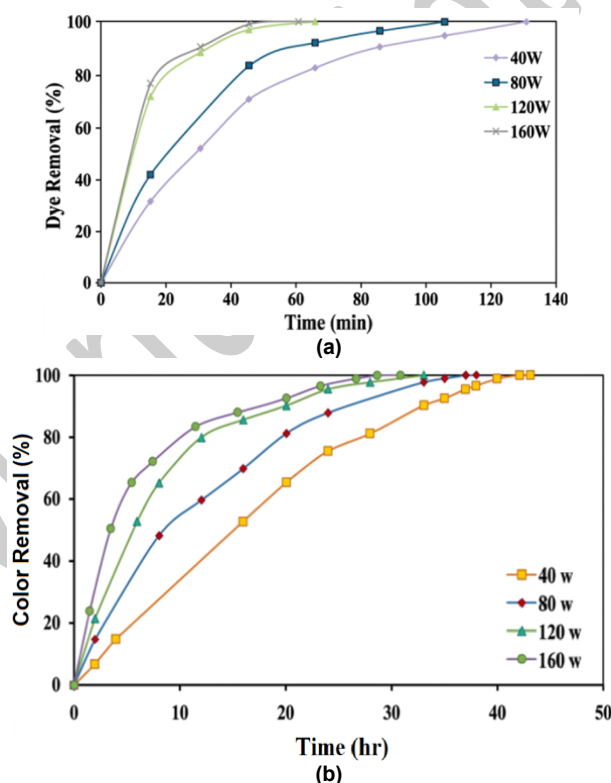


Fig. 7. (a) Hybrid and (b) photocatalytic dye removal processes efficiency in different light source powers ($[O_3]=20\text{ mg/L}$, $\text{pH}_{\text{Hybrid process}}=9$, $\text{pH}_{\text{Photocatalytic process}}=3$, $[R40]=50\text{ mg/L}$, $[\text{TiO}_2]=40\text{ gr/m}^2$).

Therefore, the hybrid process was selected as the appropriate method of dye removal and the capability of photocatalytic process was increased as the dye removed by ozonation process in system since the UV ray was completely received. All results are reported for hybrid process.

3.7. Effect of ozone concentration and injection flow rate

Considering the effect of ozone concentration and its injection flow rate on the system two cases of 12 mg/min ozone with flow rate of 0.5 L/min and 20 mg/min ozone with flow rate of 1 L/min were investigated (see Fig. 8). As shown in the figure, the hybrid process efficiency was improved as the ozone concentration and flow injection were increased. The figure also indicates that the complete removal for the ozone injection rate of 20 mg/min with flow rate of 1 L/min was achieved in less time (35 min) which was applied for the rest of the experiments.

3.8. COD removal and light penetration

Dye and COD removal efficiency in the optimum conditions of hybrid process is indicated in Fig. 9. As shown in the figure, the removal efficiency of dye (100 %) is higher than COD (18 %). This means that the chromophore functional azo dye group ($-\text{N}=\text{N}-$) were decomposed easier than the other dye structural bonds. It can also be seen that COD is increased in the first 15 minutes that can be attributed to the cleavage of cyclic compounds and non-cyclic intermediates.

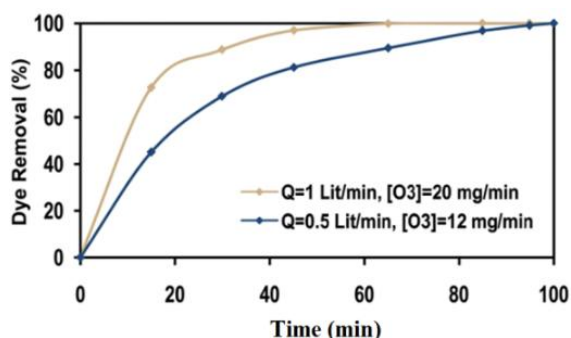


Fig. 8. Ozone injection effect on hybrid dye removal process efficiency (pH=9, [R40]=50 mg/L, [TiO₂]=40 gr/m², P_{UV-A}=120 W).

The absorbance curve variations in hybrid process are presented in Fig. 10. It can be seen that the peak of azo bond in 450-500 nm was normalized after 65 minutes while the absorbance in other wave lengths was not influenced noticeably. This means that azo bonds cleavage occurs more easily than the other bonds.

3.9. Intermediate compounds of the hybrid process

Fig. 11 shows the absorbance of different intermediates including benzene one-cyclic, di-cyclic naphthalene, and azo dye compounds in the wave lengths of 230, 485 and 310 nm, respectively. Based on the

results, the azo bonds in the solution were broken in the first hour while 3 and 7 h were needed, respectively, for cleavage of naphthalene and benzene compounds. Moreover, it can be seen that after 7 hours, all cyclic compounds were completely broken.

According to the GC-MS coupled technique after 65 minutes (Fig. 12), the compounds of 5-amino 1-naphtol, 1-naphtol compounds, phenol, aniline and phthalic acid were formed. Based on the LD₅₀ amounts (300, 275, 270, 464 and 550 mg/Kg for the mentioned organics, respectively), these compounds were more toxic than R40 with LD₅₀ of 10000 mg/Kg that was why the process was continued 7 hours to complete COD removal. This important issue has not been considered in the previous studies.

3.10. Kinetic study of the hybrid process

The kinetic relations and correlation coefficients in different experiment conditions are given in Table 3. According to the results, by increasing pH, nano TiO₂ concentration and light source power, the kinetic constant was increased whereas by increasing dye concentration, it was decreased. It can also be concluded that the pseudo first order kinetic is suitable to describe the dye removal in different conditions.

The Langmuir-Hinshelwood model is presented in the optimum conditions in Fig. 13. According to the results, there was an appropriate correlation coefficient between experimental and theoretical results. The kinetic constant and root mean square error (RMSE) were obtained 0.08 min⁻¹ and 0.3, respectively.

Table 3. The kinetic relations in the hybrid process.

Experiments conditions					Function	k, min ⁻¹	n	R ²
pH	[R40], mg/L	Ozon, mg/min	[TiO ₂], gr/m ²	Light source power, W				
3	50	20	128	120	$\log\left(\frac{-dC}{dt}\right) = -1.286 + 0.882 \cdot \log(C)$	0.051	0.882	0.993
9	50	20	128	80	$\log\left(\frac{-dC}{dt}\right) = -1.219 + 0.861 \cdot \log(C)$	0.06	0.861	0.978
9	50	20	58	120	$\log\left(\frac{-dC}{dt}\right) = -1.262 + 0.854 \cdot \log(C)$	0.055	0.854	0.987
9	35	20	128	120	$\log\left(\frac{-dC}{dt}\right) = -1.086 + 1.014 \cdot \log(C)$	0.082	1.014	0.95
7	50	20	128	120	$\log\left(\frac{-dC}{dt}\right) = -1.141 + 0.909 \cdot \log(C)$	0.072	0.909	0.95
9	50	12	128	120	$\log\left(\frac{-dC}{dt}\right) = -1.428 + 1.023 \cdot \log(C)$	0.037	1.023	0.996
9	50	20	128	40	$\log\left(\frac{-dC}{dt}\right) = -1.238 + 0.751 \cdot \log(C)$	0.057	0.751	0.971
9	50	20	191	120	$\log\left(\frac{-dC}{dt}\right) = -1.13 + 1.06 \cdot \log(C)$	0.074	1.06	0.988
9	5	20	128	120	$\log\left(\frac{-dC}{dt}\right) = -0.863 + 0.91 \cdot \log(C)$	0.137	0.91	0.989
5	50	20	128	120	$\log\left(\frac{-dC}{dt}\right) = -1.154 + 0.887 \cdot \log(C)$	0.07	0.887	0.97
9	50	20	243	120	$\log\left(\frac{-dC}{dt}\right) = -1.13 + 1.101 \cdot \log(C)$	0.074	1.101	0.983
9	50	20	128	120	$\log\left(\frac{-dC}{dt}\right) = -1.133 + 1.009 \cdot \log(C)$	0.073	1.009	0.952
9	50	20	128	160	$\log\left(\frac{-dC}{dt}\right) = -1.119 + 1.015 \cdot \log(C)$	0.076	1.015	0.95

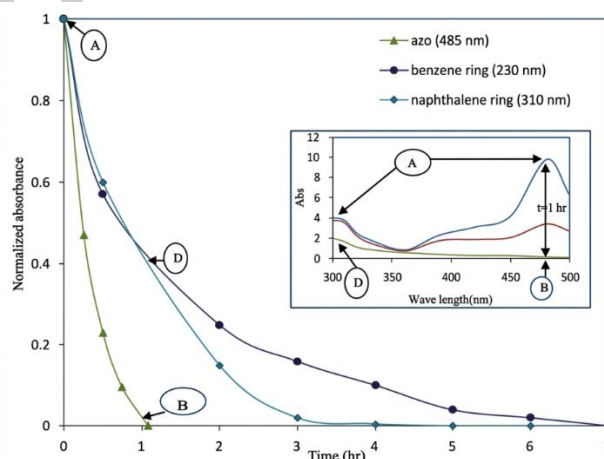


Fig. 11. Azo, naphthalene and benzene compounds absorbance in optimum conditions of hybrid process ([O₃]=20 mg/L, pH=9, [R40]=50 mg/L, [TiO₂]=40 gr/m², P_{UV-A}=120 W).

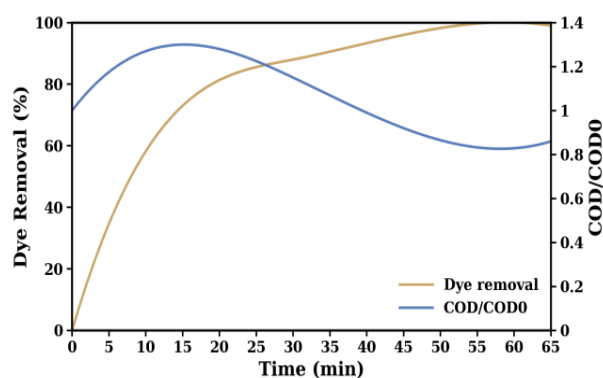


Fig. 9. Dye removal efficiency and COD removal in optimum condition of hybrid process ($[O_3]=20\text{ mg/L}$, $\text{pH}=9$, $[R40]=50\text{ mg/L}$, $[TiO_2]=40\text{ gr/m}^2$, $P_{UV-A}=120\text{ W}$).

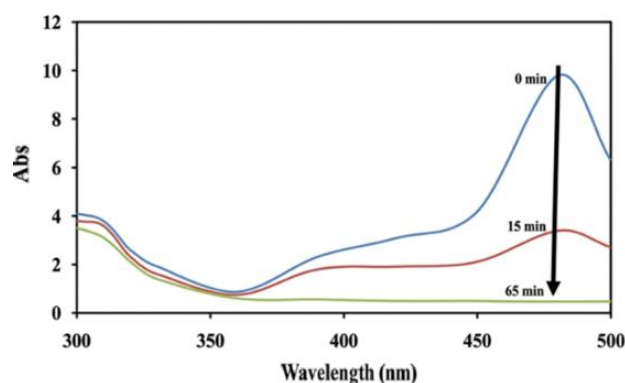


Fig. 10. Absorbance changes in optimum conditions of hybrid process ($[O_3]=20\text{ mg/L}$, $\text{pH}=9$, $[R40]=50\text{ mg/L}$, $[TiO_2]=40\text{ gr/m}^2$, $P_{UV-A}=120\text{ W}$).

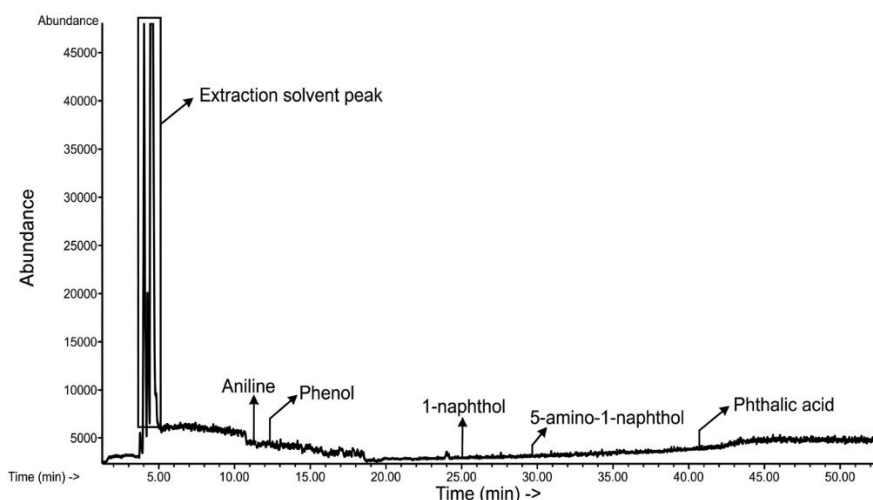


Fig. 12. Intermediates formed of dye removal in optimum hybrid process conditions ($t=65\text{ min}$).

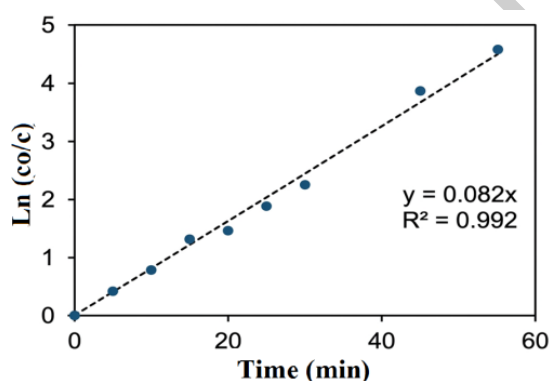


Fig. 13. Langmuir-Hinshelwood model consideration in optimum hybrid process ($[O_3]=20\text{ mg/L}$, $\text{pH}=9$, $[R40]=50\text{ mg/L}$, $[TiO_2]=40\text{ gr/m}^2$, $P_{UV-A}=120\text{ W}$).

4. Conclusions

Based on the results, improved photocatalytic (immobilized-suspended) and hybrid (Ozonation/Improved photocatalytic) methods can be used for treatment of wastewater containing azo dye of Red 40. Hybrid process had better efficiency than improved photocatalytic process and it had the best efficiency for dye concentration of 50 mg/L with pH of 9 and TiO_2 concentration of 40 g/m^3 with light source power of 120 W and ozone injection rate of 20 mg/min . In these conditions, dye and COD were removed completely after 65 min and 7 h, respectively. Based on the kinetic studies, pseudo first order model was the best model for dye degradation. Introduced hybrid process including improved immobilized-slurry photocatalytic and ozonation system not only solved recovery problem of the nano-particles but also improved the conventional process efficiency over 30 times and produced safe products to discharge. The results have been showed the advantages of hybrid process.

Author Contributions

Farhad Qaderi: Contributed to the study conception and design, conducted material preparation, data collection, analysis, and composed the original draft of the manuscript.

Amir Hossein Norouzi: Contributed to the study conception and design, data collection, and provided feedback on prior versions of the manuscript, and edited the manuscript.

Conflict of Interest

The authors have no competing interests to declare.

Data Availability Statement

Datasets employed and/or evaluated in this research are available from the corresponding author upon justifiable request.

Acknowledgment

The authors are grateful for the data support provided by Babol Noshirvani University (Babol, Iran).

References

- Amat, A.M. et al. (2005) 'Use of Ozone and/or UV in the Treatment of Effluents from Board Paper Industry', *Chemosphere*, 60(8), pp. 1111–1117. doi: <https://doi.org/10.1016/j.chemosphere.2004.12.062>
- Ayati, B. (2018) 'Modeling of a photocatalytic baffled reactor to degrade colored wastewater using response surface methodology', *Modares Civil Engineering journal*, 18(1), pp. 113-122. Available at: https://mcej.modares.ac.ir/article_12052_en.html (Accessed date: 29 July 2026).
- APHA, (2005) 'Standard Method for the Examination Water and Wastewater', USA: Washington, DC: American Public Health Association. Available at: <https://www.apha.org> (Accessed date: 12

May 2026).

- Augugliaro, V. et al. (2004) 'Photocatalytic Oxidation of Acetonitrile in Aqueous Suspension of Titanium Dioxide Irradiated by Sun Light', *Advances in Environmental Research*, 8(3-4), pp. 329–335. doi: [https://doi.org/10.1016/S1093-0191\(02\)00106-5](https://doi.org/10.1016/S1093-0191(02)00106-5)
- Azizi, M., Ebadi, T. and Qaderi, F. (2022) 'A novel method of co-doping TiO₂ with carbon and boron for enhancing photocatalytic degradation of 4-nitrophenol', *International Journal of Environmental Science and Technology*, 19, pp. 2619–2634. doi: <https://doi.org/10.1007/s13762-021-03386-z>
- Barka, N. et al. (2008) 'Photocatalytic Degradation of Indigo Carmine in Aqueous Solution by TiO₂-Coated Non Woven Fibres', *Journal of Hazardous Materials*, 152(3), pp. 1054–1059. doi: <https://doi.org/10.1016/j.jhazmat.2007.07.080>
- Behnajady, M. A. et al. (2009) 'Design Equation with Mathematical Kinetic Modeling for Photo Oxidative Degradation of C.I. Acid Orange 7 in an Annular Continuous-Flow Photoreactor', *Journal of Hazardous Materials*, 165(1-3), pp. 168–173. doi: <https://doi.org/10.1016/j.jhazmat.2008.09.095>
- Chiou, C.H., Wu, C.Y. and Juang, R.S. (2008) 'Influence of operating parameters on photocatalytic degradation of phenol in UV/TiO₂ process', *Chemical Engineering Journal*, 139(2), pp. 322–329. doi: <https://doi.org/10.1016/j.cej.2007.08.002>
- Chou, M.S. and Chang, K.L. (2007) 'Decomposition of Aqueous 2, 2, 3, 3-Tetra-Fluoro-Propanol by UV/O₃ Process', *Journal of Environmental Engineering*, 133(10), pp. 979–986. doi: [https://doi.org/10.1061/\(ASCE\)0733-9372\(2007\)133:10\(979\)](https://doi.org/10.1061/(ASCE)0733-9372(2007)133:10(979))
- Chu, L.B. et al. (2008) 'Enhanced Treatment of Practical Textile Wastewater by Micro Bubble Ozonation', *Process Safety and Environmental Protection*, 86(5), pp. 389–393. doi: <https://doi.org/10.1016/j.psep.2008.02.005>
- Delnavaz, M. et al. (2012) 'Kinetics Study of Photocatalytic Process for Treatment of Phenolic Wastewater by TiO₂ Nano Powder Immobilized on Concrete Surfaces', *Toxicological and Environmental Chemistry*, 94(6), pp. 1086–1098. doi: <https://doi.org/10.1080/02772248.2012.688331>
- Delnavaz, M. et al. (2011) 'Optimization of Photo-Catalytic Process by TiO₂ Nano Powder Immobilized on Concrete Surface for Treatment of Phenolic Wastewater', *Environmental Engineering and Management Journal*, 10(10), pp. 1459–1466. doi: <https://doi.org/10.30638/eemj.2011.206>
- Delnavaz, M., Ayati, B. and Ganjidoust, H. (2010) 'Prediction of moving bed biofilm reactor (MBBR) performance for the treatment of aniline using artificial neural networks (ANN)', *Journal of Hazardous Materials*, 179(1-3), pp. 769–775. doi: <https://doi.org/10.1016/j.jhazmat.2010.03.069>
- El-Ashtoukhy, E.S.Z. (2013) 'Removal of indigo carmine dye from synthetic wastewater by electrochemical oxidation in a new cell with horizontally oriented electrodes', *International Journal of Electrochemical Science*, 8(1), pp. 846–858. doi: [https://doi.org/10.1016/S1452-3981\(23\)14062-4](https://doi.org/10.1016/S1452-3981(23)14062-4)
- Farzinfar, B. and Qaderi, F. (2022) 'Synergistic degradation of aqueous p-nitrophenol using DBD plasma combined with ZnO photocatalyst', *Process Safety and Environmental Protection*, 168, pp. 907–917. doi: <https://doi.org/10.1016/j.psep.2022.10.060>
- Fujishima, A., Rao, T.N. and Tryk, D.A. (2000) 'Titanium dioxide photocatalysis', *Journal of Photochemistry and Photobiology C: Photochemistry Reviews*, 1(1), pp. 1–21. doi: [https://doi.org/10.1016/S1389-5567\(00\)00002-2](https://doi.org/10.1016/S1389-5567(00)00002-2)
- Ghasemi, A.H., Zoqi, M.J., and Zanganeh Ranjbar, P. (2024) 'Enhanced photocatalytic degradation of methylene blue using a novel counter-rotating disc reactor', *Frontiers in Chemistry*, 12, doi: <https://doi.org/10.3389/fchem.2024.1335180>
- Gholamian, A. et al. (2025) 'Optimization of key parameters affecting the treatment efficiency of pharmaceutical wastewater containing ciprofloxacin using TiO₂ nanoparticle photocatalytic process', *Environmental Engineering and Sustainable Development Journal*, 1(1), pp. 31–38. Available at: https://eesd.modares.ac.ir/?lang=en&keywords=qaderi&_action=article (Accessed date: 8 May 2026).
- Gomes, A.C., Nunes, J.C. and Simões, R.M.S. (2010) 'Determination of fast ozone oxidation rate for textile dyes by using a continuous quench-flow system', *Journal of Hazardous Materials*, 178(1-3), pp. 57–65. doi: <https://doi.org/10.1016/j.jhazmat.2010.01.043>
- Hu, T.L., Lin, C. and Chiang, K.Y. (2004) 'Simulations of organic carbon transformation in dye wastewater and treatment suggestions', *Advances in Environmental Research*, 8(3-4), pp. 493–500. doi: [https://doi.org/10.1016/S1093-0191\(02\)00147-8](https://doi.org/10.1016/S1093-0191(02)00147-8)
- Kartal, O.E., Erol, M. and Oguz, H. (2001), 'Photocatalytic destruction of phenol by TiO₂ powders', *Chemical Engineering Technology*, 24(6), pp. 645–649. doi: [https://doi.org/10.1002/1521-4125\(200106\)24:6%3C645::AID-CEAT645%3E3.0.CO;2-L](https://doi.org/10.1002/1521-4125(200106)24:6%3C645::AID-CEAT645%3E3.0.CO;2-L)
- Kasperchik, V.P., Yaskevich, A.L. and Bil'dyukovich A.V. (2012) 'Wastewater Treatment for Removal of Dyes by Coagulation and Membrane Processes', *Petroleum Chemistry*, 52(7), pp. 545–556. doi: <https://doi.org/10.1134/S0965544112070079>
- Khani, A. et al. (2013) 'Enhancing purification of an azo dye solution in nanosized zero-valent iron-ZnO photocatalyst system using subsequent semi batch packed-bed reactor', *Turkish Journal Engineering Environmental Sciences*, 37, pp. 91–99. doi: <https://doi.org/10.3906/muh-1201-6>
- khourshidi, A. and Qaderi, F. (2023) 'Optimization of p-nitrophenol-contaminated water by non-thermal plasma technology and ozonation by response surface method', *Modares Civil Engineering journal*, 23(5), pp. 179–193. doi: <https://doi.org/10.22034/23.5.13>
- Lackey, L.W., Mines Jr., R.O. and McCreanor, P.T. (2006) 'Ozonation of acid yellow 17 dye in a semi-batch bubble column', *Journal of Hazardous Materials*, 138(2), pp. 357–362. doi: <https://doi.org/10.1016/j.jhazmat.2006.05.116>
- Lu, X. et al. (2009) 'Treatment of wastewater containing azo dye reactive Brilliant Red X-3B using sequential ozonation and upflow biological aerated filter process', *Journal of Hazardous Materials*, 161(1), pp. 241–245. doi: <https://doi.org/10.1016/j.jhazmat.2008.03.077>
- Mirkhani, V. et al. (2009) 'Photocatalytic degradation of azo dyes catalyzed by ag doped TiO₂ photocatalyst', *Journal of the Iranian Chemical Society*, 6(3), pp. 578–587. doi: <https://doi.org/10.1007/BF03246537>
- Moradianfard, Sh., Ghorbani, J. and Bakhshy, K. (2013) 'Color and COD removal from wastewater containing acrylic water base color using fenton's oxidation process', *International Journal of Advanced Biological and Biomedical Research*, 1(1), pp. 45–52. Available at: https://www.ijabbr.com/article_6570.html (Accessed date: 14 January 2026).
- Papadam, T. et al. (2007) 'Photocatalytic transformation of acid orange 20 and Cr (VI) in aqueous TiO₂ suspensions', *Journal of Photochemistry and Photobiology A: Chemistry*, 186(2-3), pp. 308–315. doi: <https://doi.org/10.1016/j.jphotochem.2006.08.023>
- Peng, Y. et al. (2008) 'NaNO₂/FeCl₃ catalyzed wet oxidation of the azo dye Acid Orange 7', *Chemosphere*, 71(5), pp. 990–997. doi: <https://doi.org/10.1016/j.chemosphere.2007.10.065>
- Prahba, I. and Lathasree, S. (2013) 'Photocatalytic performance of nanocatalyst for the effective removal of dye in the wastewater', *Chemical science transactions*, doi: <https://doi.org/10.7598/cst2013.021>
- Pratap Reddy, M. et al. (2004) 'An integrated approach of solar photocatalytic and biological treatment of N-containing organic compounds in wastewater', *Toxicological and Environmental Chemistry*, 86(3), pp. 127–140. doi: <https://doi.org/10.1080/02772240410001665526>
- Qaderi, F. and Abdolalian, S. (2022) 'Treatment of petroleum wastewater using solar power-based photocatalysis', *Petroleum Industry Wastewater*, pp. 161–170, doi: <https://doi.org/10.1016/B978-0-323-85884-7.00009-6>
- Qaderi, F., Ayati, B. and Ganjidoust, H. (2011) 'Role of moving bed biofilm reactor and sequencing batch reactor in biological degradation of formaldehyde wastewater', *Journal of Environmental Health Science and Engineering*, 8(4), pp. 295–306. Available at: <https://ijehse.tums.ac.ir/index.php/jehse/article/view/322/0> (Accessed date: 12 May 2026).
- Qaderi, F., Banisheikholeslami, A. and Tamadoni, A., (2025) 'Intelligent models as novel tools for optimizing ultrasonication-ozonation technique in PAH-contaminated soil remediation', *Journal of Environmental Health Science and Engineering*, doi:

<https://doi.org/10.21203/rs.3.rs-4061528/v1>

- Qaderi, F., Tamadoni, A. and Banisheikhosslami, A. (2024) 'Cost estimation for application of ultrasonication–ozonation hybrid process in remediation of PAH-contaminated soil', *Environment Development and Sustainability*, 26, pp. 12441–12466. doi: <https://doi.org/10.1007/s10668-023-03828-3>
- Rabieian, M. and Qaderi, F. (2024) 'Optimizing hybrid photocatalytic-ozonation for offshore produced water treatment', *Journal of Mining and Environment*, 15(1), pp. 239-259. doi: <https://doi.org/10.22044/jme.2023.13081.2376>
- Ranjbar, P. Z., Ayati, B. and Ganjidoust, H. (2022) 'Textile dye degradation in a novel photocatalytic baffled reactor using immobilised TiO₂ nanoparticles', *International Journal of Environment and Waste Management*, 29(3), pp. 241-261. doi: <https://doi.org/10.1504/IJEW.2022.122677>
- Ranjbar, P.Z., Ayati, B., and Ganjidoust, H. (2019) 'Kinetic study on photocatalytic degradation of Acid Orange 52 in a baffled reactor using TiO₂ nanoparticles', *Journal of Environmental Sciences*, 79, pp. 213-224. doi: <https://doi.org/10.1016/j.jes.2018.06.012>
- Somensi, C.A. *et al.* (2010) 'Use of ozone in a Pilot-scale plant for textile wastewater pre-treatment: physico-chemical efficiency, degradation by-products identification and environmental toxicity of treated wastewater', *Journal of Hazardous Materials*, 175(1-3), pp. 235–240. doi: <https://doi.org/10.1016/j.jhazmat.2009.09.154>
- Tamadoni, A. and Qaderi, F. (2019) 'Optimization of soil remediation by ozonation for PAHs contaminated soils', *The Journal of the International Ozone Association*, 41(5), pp. 454-472. doi: <https://doi.org/10.1080/01919512.2019.1615865>
- Turhan, K. and Turgut, Z. (2009) 'Decolorization of direct dye in textile wastewater by ozonation in a semi-batch bubble column reactor', *Desalination*, 242(1-3), pp. 256–263. doi: <https://doi.org/10.1016/j.desal.2008.05.005>
- Weng, M., Zhou, Z. and Zhang, Q. (2013) 'Electrochemical degradation of typical dyeing wastewater in aqueous solution: performance and mechanism', *International Journal of Electrochemical Science*, 8(1), pp. 290-296. doi: [https://doi.org/10.1016/S1452-3981\(23\)14020-X](https://doi.org/10.1016/S1452-3981(23)14020-X)
- Wu, Y. *et al.* (2010) 'Properties of carbon and iron modified TiO₂ photocatalyst synthesized at low temperature and photodegradation of acid orange 7 under visible light', *Applied Surface Science*, 256(13), pp. 4260–4268. doi: <https://doi.org/10.1016/j.apsusc.2010.02.012>
- Zhu, X., Zhang, J. and Chen, F. (2010) 'Hydrothermal synthesis of nanostructures Bi₁₂TiO₂₀ and their photocatalytic activity on acid orange 7 under visible light', *Chemosphere*, 78(11), pp. 1350–1355. doi: <https://doi.org/10.1016/j.chemosphere.2010.01.002>
- Zuorroa, A. and Lavecchia, R. (2013) 'Evaluation of UV/H₂O₂ advanced oxidation process (AOP) for the degradation of diazo dye reactive green 19 in aqueous solution', *Desalination and Water Treatment*, 52(7-9), pp. 1571-1577. doi: <https://doi.org/10.1080/19443994.2013.787553>